



# Reducing Reactive-N Loss from Fertigation: High-Frequency Application and Fertilizer Selection

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## Introduction

High-Frequency Low-Nitrogen fertigation (HFLN), also called "split-application," is a little-investigated and simple approach which may reduce nitrate ( $\text{NO}_3^-$ ) leaching. It may also reduce emissions of the important greenhouse gas nitrous oxide ( $\text{N}_2\text{O}$ ) in some systems, depending upon the N fertilizer used, and upon soil microbial community effects. Choice of N fertilizer can have important effects on the location and magnitude of  $\text{N}_2\text{O}$  production (Smith *et al.*, 1997), as well as on the microbial processes which applied N will undergo. Information on the quantity of  $\text{NH}_4^+$ ,  $\text{NO}_3^-$  and  $\text{N}_2\text{O}$  at different depths in the soil profile following fertigation is essential to explaining any differences seen in  $\text{NO}_3^-$  loss or  $\text{N}_2\text{O}$  emission when comparing N-fertilizers and N-fertigation systems.

## Treatments

1. **UAN-Standard:** UAN applied over 4 fertigations, 300 lbs N/acre
2. **UAN-HFLN:** UAN applied over 20 fertigations, 300 lbs N/acre
3.  **$\text{NO}_3^-$ -HFLN:**  $\text{Ca}(\text{NO}_3)_2$  (61%) +  $\text{KNO}_3$  (39%) applied over 20 fertigations, 300 lbs N/acre

## Hypotheses

HFLN was expected to reduce  $\text{NO}_3^-$  leaching and total  $\text{N}_2\text{O}$  emissions. Within HFLN, we expected a high- $\text{NO}_3^-$  treatment to yield lowest  $\text{N}_2\text{O}$  emissions. **Determining factors were expected to include** N concentration in soil solution, species of N, conditioning of soil microbial community, and location of applied N in the soil profile.

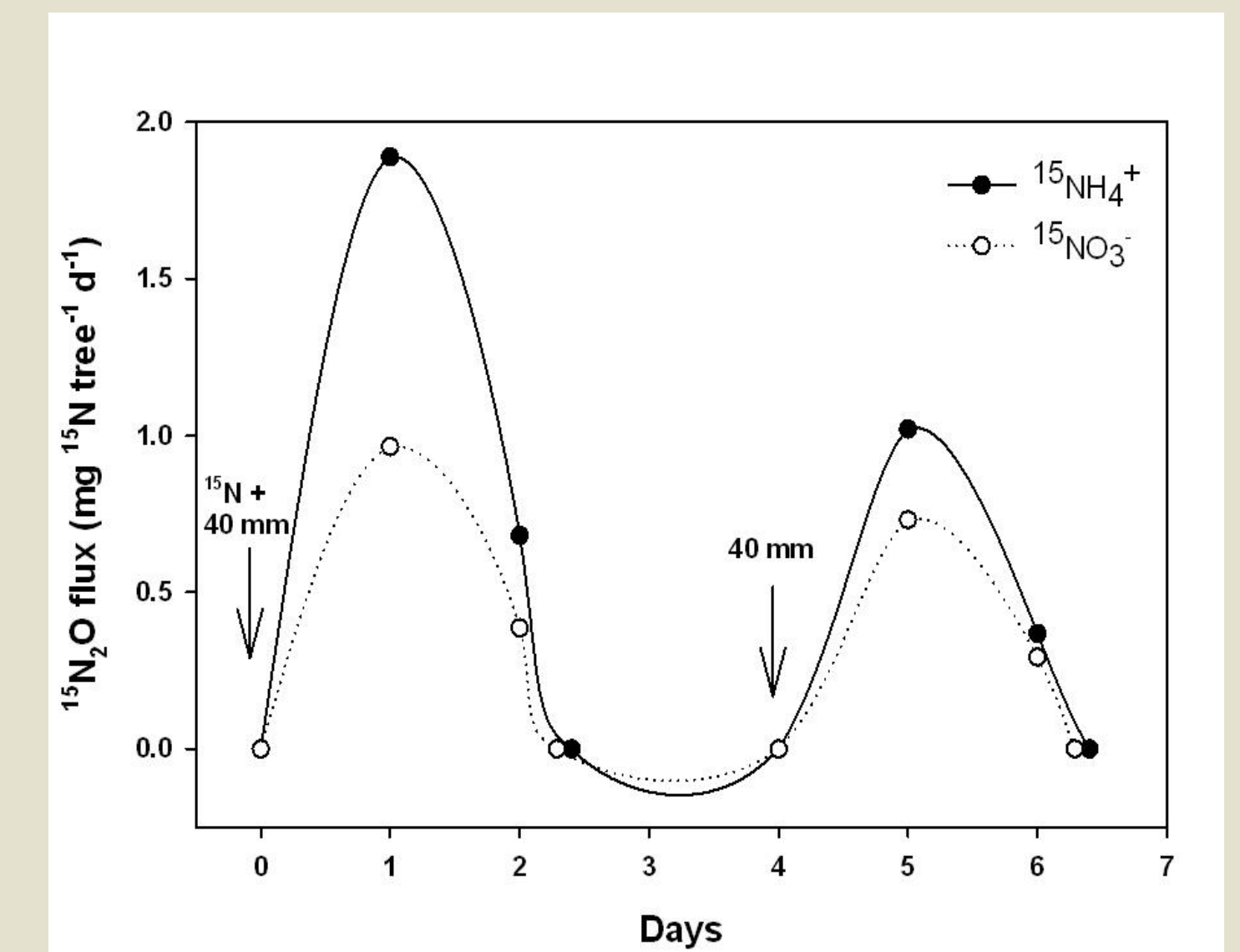
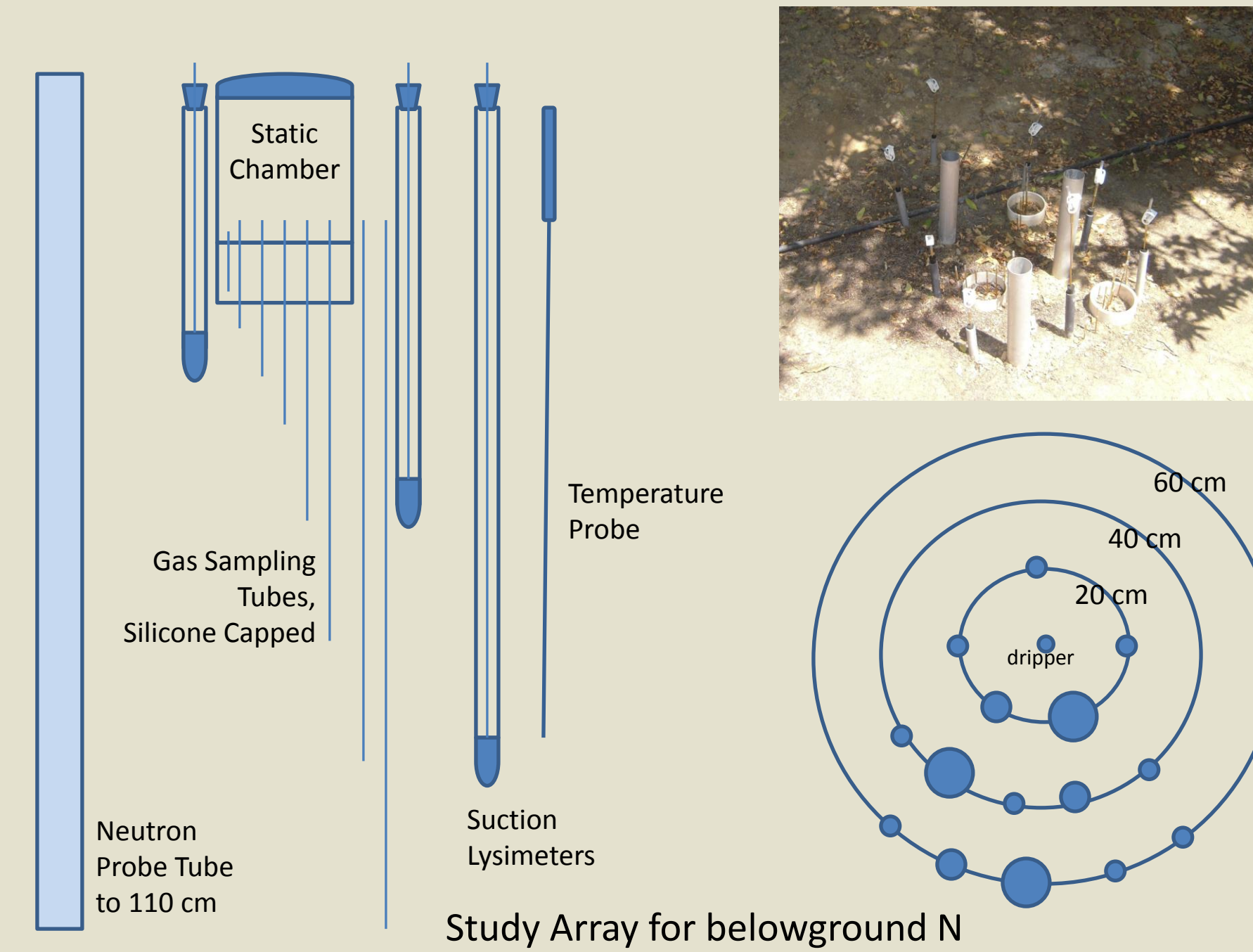
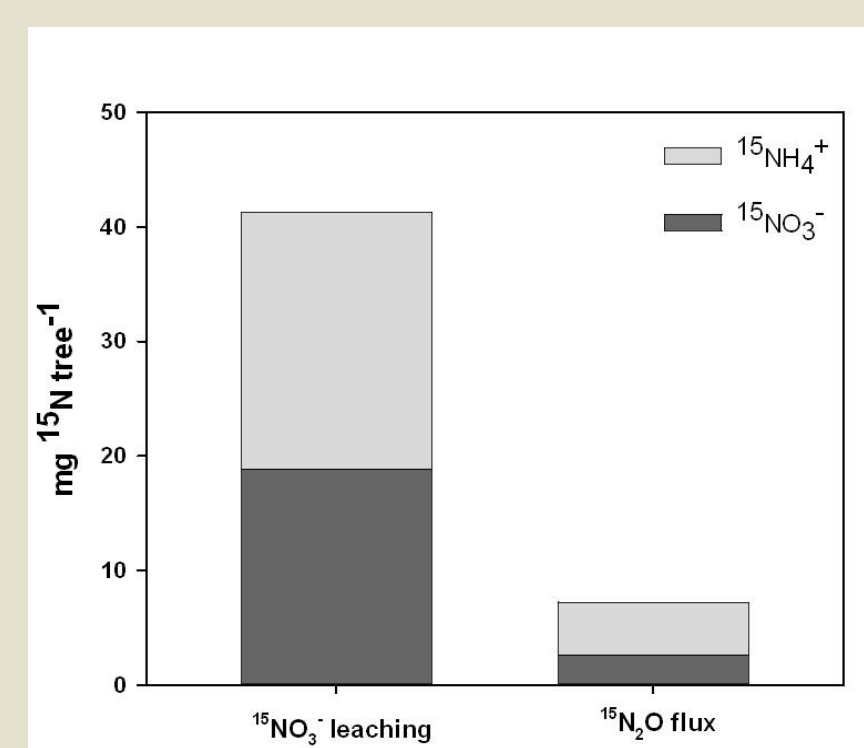
## Preliminary Results

1.  $\text{NO}_3^-$  leaching results do not appear likely to show differences. However, the higher quantities of  $\text{N}_2\text{O}$  found at 80 cm depth under Standard fertigation suggest greater N penetration into the soil profile under Standard.
2. The first year's data showed **slightly lower yield under HFLN**, although differences were not significant. HFLN scheduling may require modification.
3. **HFLN with UAN has little to no effect on  $\text{N}_2\text{O}$  emissions** vs. standard fertigation strategies.
4. **HFLN with high- $\text{NO}_3^-$  fertilizers lowers  $\text{N}_2\text{O}$  emissions significantly compared to HFLN-UAN.** HFLN may be a good option using high- $\text{NO}_3^-$  fertilizer formulations, with lower N concentrations reducing  $\text{NO}_3^-$  leaching risks.
5. Patterns of  $\text{N}_2\text{O}$  production in the soil profile clearly show the **spatial separation of processes affecting different N fertilizer sources**. Very high amounts of  $\text{N}_2\text{O}$  are recorded in the soil profile, up to 1000x higher than ambient. Most  $\text{N}_2\text{O}$  appears to be consumed within the soil rather than emitted from the surface.

## Preliminary $\text{N}_2\text{O}$ Emissions

Fertigation Strategy	Fertilizer	Emission Factor	Annual $\text{N}_2\text{O}$ Emissions	Sig. ( $p < .05$ )
Standard	UAN	1.08%	3.24 kg $\text{N}_2\text{O}$ -N $\text{ha}^{-1} \text{yr}^{-1}$ a	
HFLN	UAN	0.90%	2.70 kg $\text{N}_2\text{O}$ -N $\text{ha}^{-1} \text{yr}^{-1}$ a	
HFLN	$\text{Ca}(\text{NO}_3)_2$	0.49%	1.47 kg $\text{N}_2\text{O}$ -N $\text{ha}^{-1} \text{yr}^{-1}$ b	

## Preliminary Comparison of $\text{NO}_3^-$ Loss and $\text{N}_2\text{O}$ Emissions



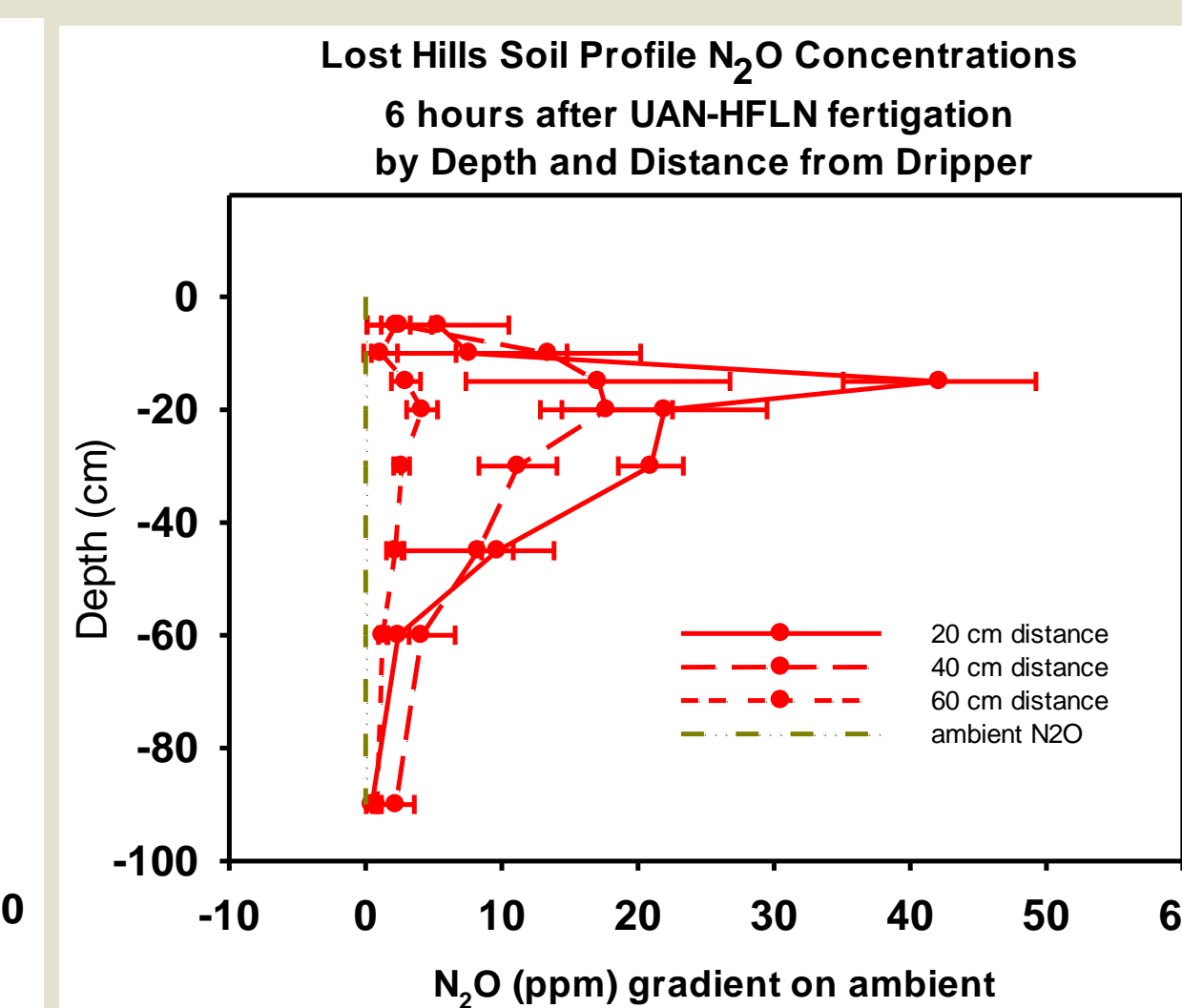
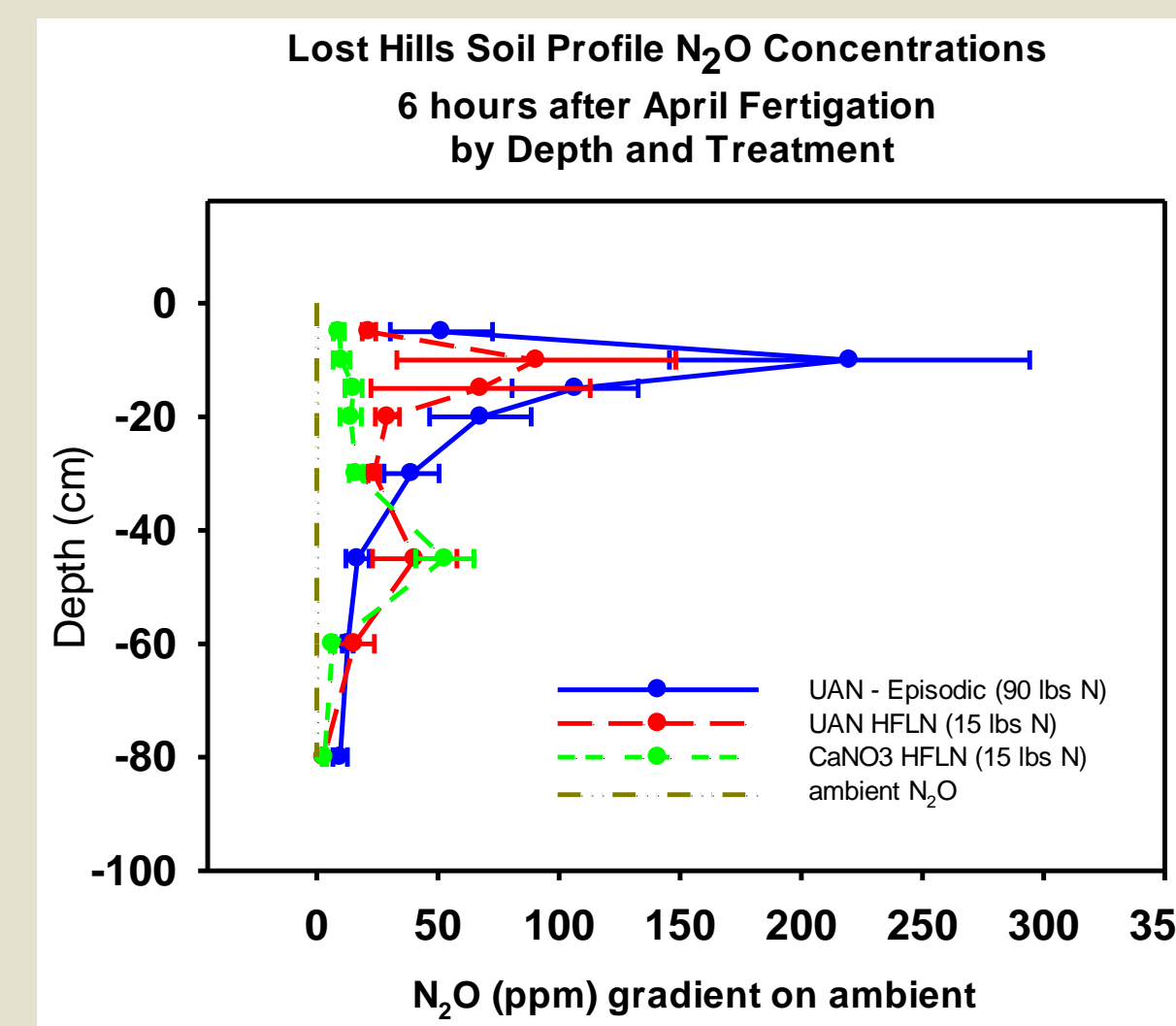
## Origin of $\text{N}_2\text{O}$ from UAN application

Comparing emissions results from the above field  $^{15}\text{N}$  tracer study with soil gas data on the left, it is suggested that nitrification of ammonium applied as UAN in upper levels of the soil contributes to the higher emissions seen from UAN than from  $\text{Ca}(\text{NO}_3)_2$ . Further research remains to be done on this question.

Net  $\text{N}_2\text{O}$  Production/Consumption Rate per soil layer, calculated using points of profile gas concentration

$$\alpha_i = \frac{q_i - q_{i+1}}{Z_i + Z_{i+1} - \frac{Z_i + Z_{i-1}}{2}} + \frac{dc_i}{dt} V_{Ai}$$

Cumulative Flux ( $\text{g} \cdot \text{tree}^{-1}$ )	0 - 10 cm		0 - 50 cm	
	Mean	SE	Mean	SE
Gross mineralization (m)	8.70	3.15	11.4	3.80
$\text{NH}_4^+$ consumption ( $c_{\text{NH}_4}$ )	12.3	2.38	15.1	2.93
Gross nitrification (n)	5.20	0.46	12.3	1.74
$\text{NO}_3^-$ consumption ( $c_{\text{NO}_3}$ )	6.63	0.34	15.1	4.50
Gross N immobilization (i)	13.3	2.24	16.7	5.52



## $\text{N}_2\text{O}$ Production/Consumption in Soil

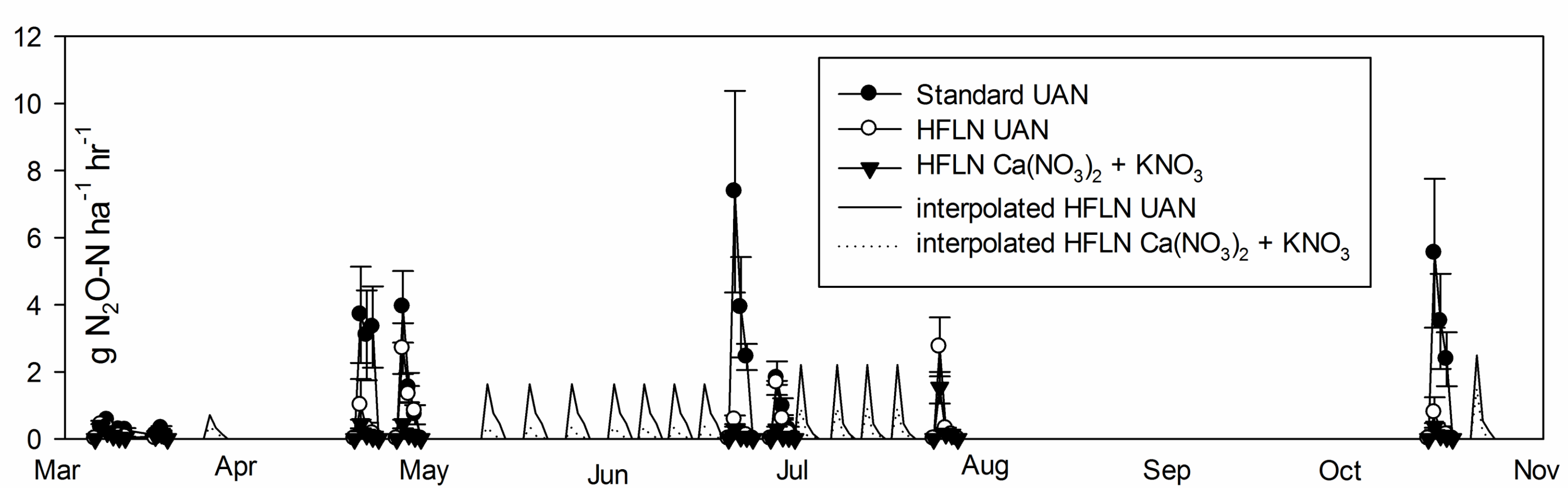
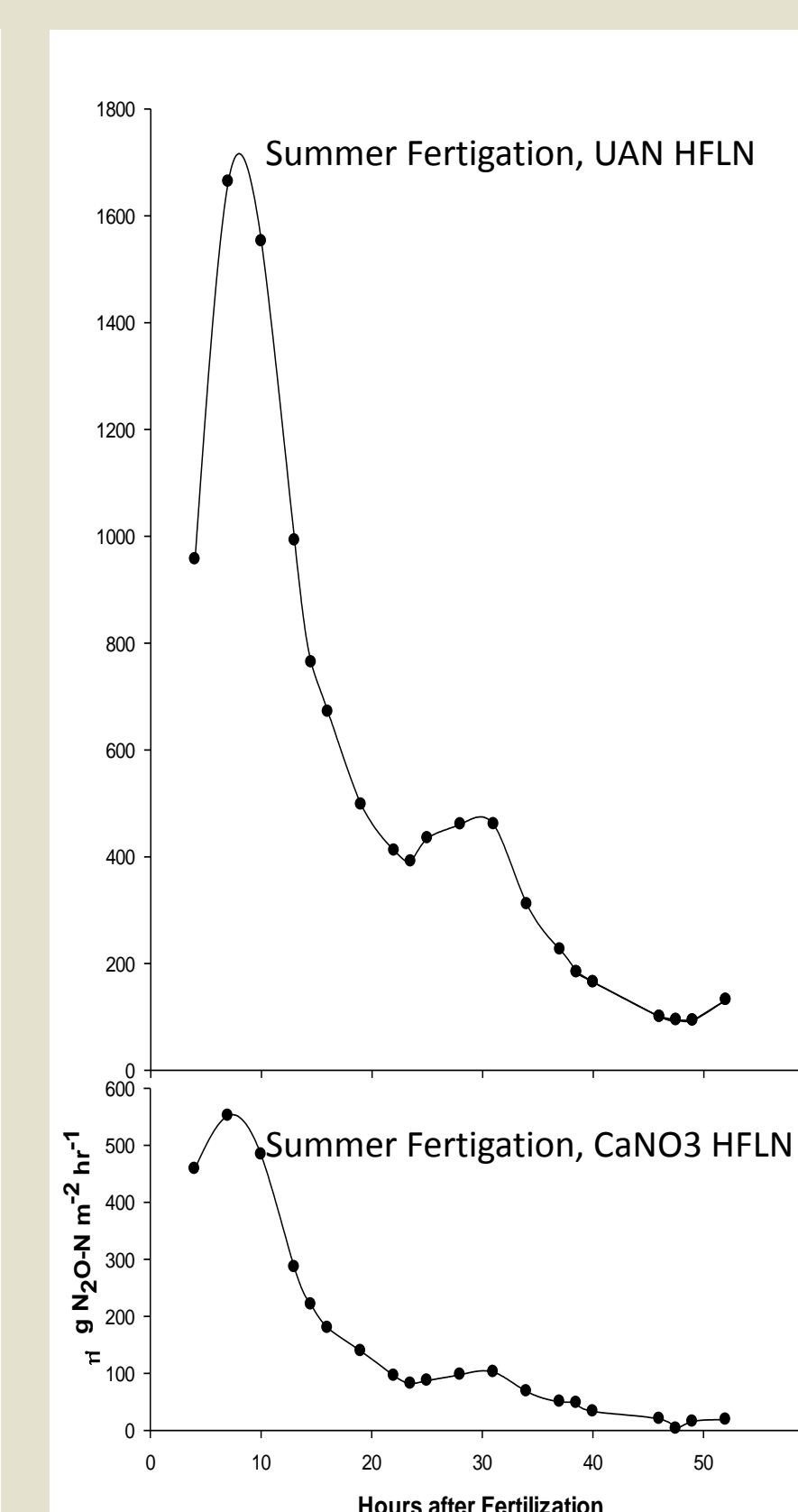
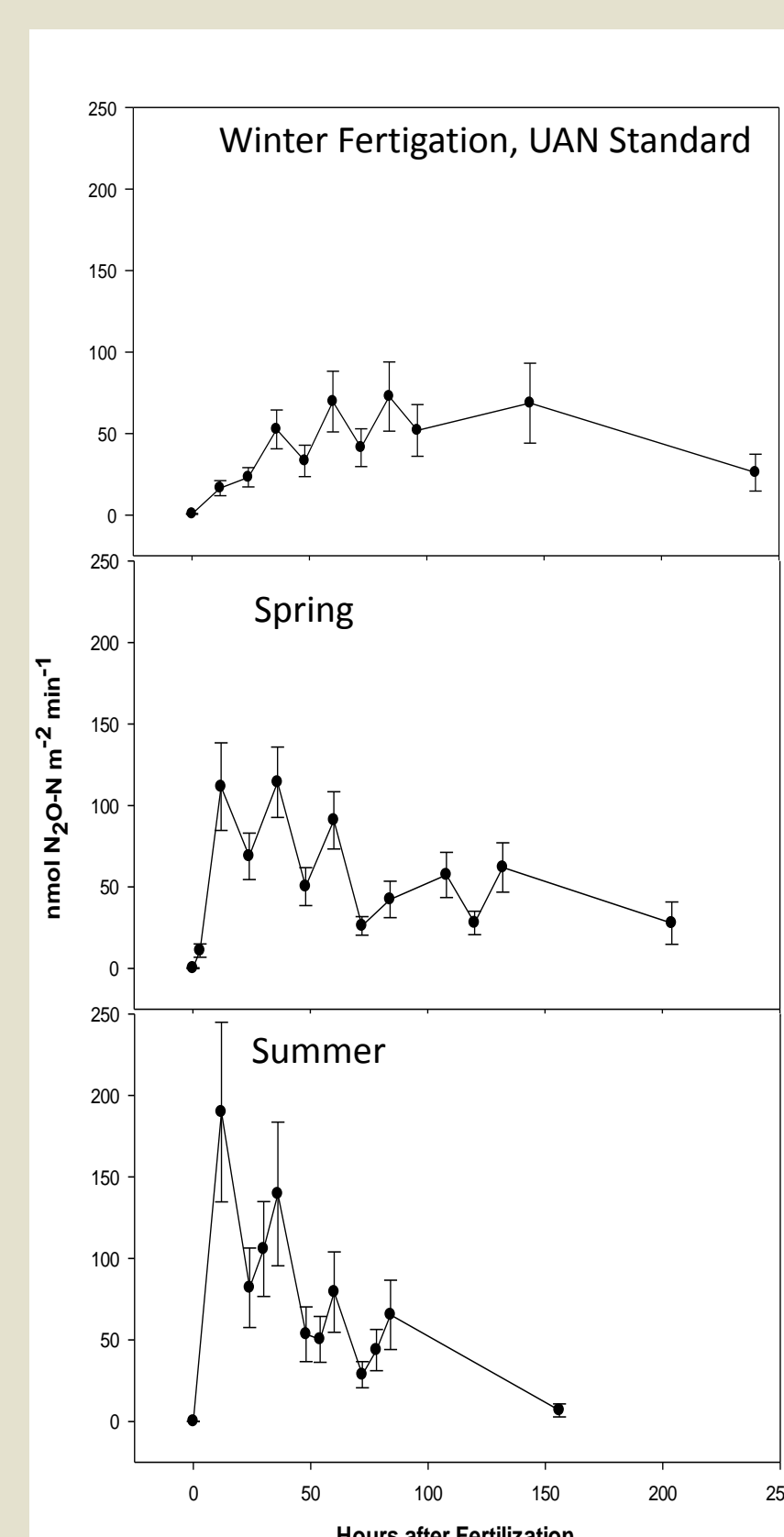
Combining results like those above with water-filled pore space data (WFPS) can reveal rates of net  $\text{N}_2\text{O}$  production and consumption, using the 1-dimensional model to right, above (Yoh, 1997).

Studies of  $^{15}\text{N}$  fate in this soil (right) show the effects of soil depth on the dominant processes affecting applied Nitrogen.

Combined with soil parameters, such data may facilitate the development of  $\text{N}_2\text{O}$  predictions using hydrological models.

## Diurnal $\text{N}_2\text{O}$ Emissions after Fertigation

Careful diurnal monitoring is necessary to quantify  $\text{N}_2\text{O}$  emissions following applications of fertilizer N. Temperature differences cause varying patterns by season, left and below.



## Methods

$\text{N}_2\text{O}$  emissions and  $\text{NO}_3^-$  losses are being assessed in a San Joaquin Valley almond orchard (*Prunus dulcis*) in Belridge, California under drip fertigation laid out in an RCB design with 5 blocks. Yield is being evaluated over 5 years. The soil is a calcareous sandy loam (pH 8.3) with two clay layers (50-80 cm, approx. 30% clay; 180-250 cm approx. 45% clay).  $\text{N}_2\text{O}$  emissions are evaluated using static chambers with 2, 10-minute fluxes. Evaluations take place at each Standard fertigation, 4x/year (February, April, June, September). They include emissions from the first 3-5 days following fertigation, as well as the 3 days following the next irrigation. Chambers are placed at 4 distances from the dripper (0, 20, 50, 90 cm), which are used for emissions estimates assuming radial symmetry. Treatments are compared as  $\text{N}_2\text{O}$  emitted/N applied. At 20 cm from the dripper, the location of highest  $\text{N}_2\text{O}$  efflux, 1/8" brass tubes sample soil gas at 5,10,15,20,30,45,60 and 80 cm deep. Suction lysimeters sample soil solution for  $\text{NH}_4^+$  and  $\text{NO}_3^-$  at 15, 30 and 60 cm. A Campbell Hydroprobe is used to evaluate water-filled pore-space at intervals from 15 to 90 cm. A Decagon ECH<sub>2</sub>O probe is used to estimate surface WFPS.  $\text{NO}_3^-$  losses are estimated using suction lysimeters at 250 cm compared to results from tensiometers at 220 cm and 280 cm to estimate infiltration rates.

## References

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